

1 **BIODIVERSITY REVIEW**

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3 **Alien freshwater fishes in Brazilian aquaculture and sport fishing: reviewing**
4 **their economic benefits and risks to native fish diversity**

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14 **Running title:** Risks from alien freshwater fishes in Brazil

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25ABSTRACT

26

27**Aim:** To review the use and value of alien freshwater fishes in Brazilian aquaculture
28and recreational angling, their ecological risks in the environment and the empirical
29evidence on their impacts on native fish diversity.

30**Location:** Brazil and the Neotropical Region.

31**Methods:** Literature review, analysis of relevant databases on aquaculture and
32freshwater fish introductions, and use of the Fish Invasiveness Scoring Kit (FISK).

33**Results:** The use of alien freshwater fishes in aquaculture and angling in Brazil
34specifically, and the Neotropical Region generally, has never been subject to
35ecological risk assessment processes. Today, alien fishes comprise a substantial
36proportion of total aquaculture production and value in Brazil, contributing over \$US
37250 million in 2008 (63 % of the total value of freshwater aquaculture). The alien
38fishes that are used most prominently are *Cyprinus carpio* and *Oreochromis niloticus*;
39ecological risk assessment and literature review suggest both species are potentially
40highly invasive and may affect the ecological integrity of invaded water bodies.
41Peacock basses (*Cichla* species) have been introduced from Amazonia into other
42Brazilian river basins for creation of recreational fisheries. Introductions generally
43result in establishment and invasion, and significant and rapid declines in native fish
44diversity have been subsequently reported in a number of studies.

45**Main conclusions:** Alien freshwater fishes in Brazilian aquaculture and sport fishing
46support ecosystem services of economic value, but the principal species used
47represent high ecological risks, with empirical studies revealing deleterious impacts
48on native fishes, including extinctions. It is thus recommended that risk assessment
49and risk management processes on alien fishes are implemented in Brazil and the

50Neotropical Region to minimise the threat of further harmful introductions, in
51conjunction with the continued development and support of aquaculture systems
52based on indigenous fishes.

53

54**Keywords:** *Cichla kelberi*, *Oreochromis niloticus*, risk assessment, invasion, non-
55indigenous.

56

57INTRODUCTION

58

59The ecosystem services of freshwater aquaculture and recreational angling have the
60capacity to generate significant economic returns, support recreational activities with
61substantial societal benefits, and play an important role in global food production
62(Pitcher, 1999; Ross *et al.*, 2008). For example, global freshwater fish aquaculture
63production in 2009 exceeded 73 million tonnes, worth approximately \$US 61 billion
64(FAO, 2011a). Although similar global figures are not available for recreational
65angling, national figures suggest it supports significant regional and national
66economies, with an estimated combined annual expenditure in ten countries in
67Western Europe during the 1990s of \$US 10 billion (Cowx, 1998). Both aquaculture
68and angling have a tendency to rely on alien fishes to enhance production and/ or
69fishery performance (Casal, 2006; Gozlan, 2008; Gozlan *et al.*, 2010). Globally,
70species such as rainbow trout *Oncorhynchus mykiss*, common carp *Cyprinus carpio*
71and Nile tilapia *Oreochromis niloticus* are now present in many countries having been
72introduced for these purposes (Casal, 2006; Zambrano *et al.*, 2006; Vitule *et al.*,
732009). Indeed, fishes are among the most introduced group of aquatic animals in the

74world (Gozlan, 2008), with aquaculture (51 % of all introductions) and angling (12 %
75of all introductions) providing major introduction pathways (Gozlan *et al.*, 2010).

76

77 The fish diversity of Brazilian freshwaters has high global significance; by 2003, at
78least 2122 fishes were catalogued (21 % of the world list; Agostinho *et al.* 2005).
79Moreover, fish assemblages in isolated river basins tend to comprise of high
80proportions of endemic fishes (Agostinho *et al.*, 2005) and it has been argued that as
81much as 40 % of Neotropical freshwater fishes have yet to be described (Reis *et al.*,
822003). A proportion of these fishes are, however, threatened; Agostinho *et al.* (2005)
83discussed factors that were causing threats to 134 species and these included river
84impoundments, fishery exploitation and introductions of alien species. These were
85most apparent in the more developed regions of Brazil, particularly those in the south
86and southeast (Agostinho *et al.*, 2005). Reservoirs have been built in the majority of
87Brazilian hydrographic basins and these have profoundly influenced the composition
88and structure of the fish assemblages (Agostinho *et al.*, 2007, 2008). For example,
89comparisons of assemblage pre- and post dam construction generally show substantial
90losses of migratory species with replacement by sedentary species (Gubiani *et al.*,
912010). Where previously impassable barriers to fish movement have been overcome
92by water inundation then this has also enabled the movement of many fishes into
93previously isolated watersheds (Júlio Júnior *et al.*, 2009).

94

95 It is against this background of substantial environmental and habitat modification
96through impoundment that the influence of alien species in aquaculture and
97recreational angling on freshwater fish diversity in Brazil has to be assessed. The
98formation of substantial areas of lentic water has enabled the development of new

99recreational fisheries and cage aquaculture systems in what were previously lotic
100waters (Agostinho *et al.*, 2005). For example, in the upper Paraná River Basin,
101recreational fisheries have been created through introductions of peacock basses -
102fishes of the *Cichla* genus - including *Cichla kelberi*, *Cichla piquiti* and *Cichla*
103*monoculus* (Latini & Petrere, 2004; Viera *et al.*, 2009). Cage aquaculture systems
104have developed that frequently utilise species such as *Oreochromis niloticus*
105(Zambrano *et al.*, 2006; Vitule *et al.*, 2009). From a conservation perspective, the
106precautionary approach would suggest it inappropriate to utilise these alien fishes in
107this manner (Leprieur *et al.*, 2009; Vitule *et al.*, 2009). However, of the numerous
108freshwater fishes that have been introduced into areas outside of their natural range,
109establishment rates have been relatively low and few species have had measureable
110ecological impacts (Gozlan, 2008). Instead, many introduced fishes have provided
111considerable societal and economic benefits (Gozlan *et al.*, 2010). Consequently, it
112can be argued that policy formulation and management of alien fishes in the
113environment requires evidence-based science on which to base decisions, rather than
114relying on precaution (Britton *et al.*, 2011a,b).

115

116 The use of alien freshwater fishes in aquaculture and angling in Brazil specifically,
117and the Neotropical Region more generally, has previously never been subject to
118ecological risk assessment processes (Agostinho *et al.*, 2005; Darrigran *et al.*, 2011).
119Although less satisfactory, risk assessment can be used in post-introduction phases as
120this can help inform subsequent management approaches (Britton *et al.*, 2011a).
121When used in combination with published empirical studies on their impacts, *a*
122*posteriori* assessments of alien fishes can be compared with the economic benefits
123they have accrued. Thus, the aim of this review was to identify the utilisation and

124value of alien fishes in Brazilian freshwater aquaculture and recreational angling,
125utilise risk assessment techniques to predict the ecological risks of these fishes, review
126empirical studies to evaluate their actual impacts on fish diversity, and finally evaluate
127their continued use in these ecosystem services.

128

129**MATERIALS AND METHODS**

130

131Open-access and independently-collected databases were searched to identify the
132alien freshwater fishes currently present in Brazil, their introduction pathways
133(aquaculture, angling or other), and their production levels and values. The databases
134were those relating aquaculture (Food and Agriculture Organisation FAO;
135<http://www.fao.org/fi/statist/statit.asp>) and fish introductions (FishBase
136<http://www.fishbase.org>; Global Invasive Species Database
137<http://www.issg.org/database/welcome/>). All were accessed on 17/04/2011.

138

139 Assessments of the ecological risks associated with the alien fishes in Brazilian
140freshwaters were completed using 2 methods. Firstly, scores were used from the Fish
141Invasiveness Scoring Kit (FISK; CEFAS, 2011), where scores were already available
142for the majority of the alien fishes and where a score > 19 indicates a high risk of
143invasiveness (*cf.* Copp *et al.*, 2009; Britton *et al.*, 2010). Although primarily
144developed for use in the UK and Western Europe (Copp *et al.*, 2005), the toolkit has
145wider application in assessing the ecological risks associated with introduced fishes
146elsewhere (Copp *et al.*, 2009). Where scores were not available for a species, FISK
147was utilised by the authors and an assessment completed as per Copp *et al.* (2009).
148Secondly, assessments were made as per Gozlan (2010), where the likelihood of an

149alien fish invoking ecological impacts in the environment was assessed from literature
150reviews. Impact was defined as a negative consequence in a native ecosystem that is
151quantifiable and has resulted in habitat degradation, competition with native species
152for spawning ground, hybridisation threatening species integrity and/or predation on
153native species populations, resulting in their decline (Gozlan, 2010). Similar to FISK,
154for a number of the alien fishes present in Brazil, impact likelihood values (as the
155percentage of studies where impact was detected) were already present in Gozlan
156(2010); where these were absent then the reviews were completed by the authors.

157

158 Finally, reviews (case studies) on two groups of alien fishes present in Brazilian
159freshwater aquaculture and angling were developed, with the reviewed fishes those
160that were assessed as having the potential to cause ecological impacts. The reviews
161were designed to identify whether empirical evidence was available from field studies
162to corroborate the assessed risks. This then enabled the empirical information to be
163compared against as the value of these alien fishes in supporting the ecosystem
164services so that their overall benefits to Brazil could be evaluated.

165

166RESULTS

167

168Alien freshwater fishes in Brazil

169There at least 15 alien freshwater fishes present in Brazil that are associated with the
170aquaculture trade or angling (Table 1); in the case *Cichla* species, these have been
171introduced into impoundments in the south and southeast of the country from
172Amazonia for creating angling opportunities and so represent a translocation within
173the country rather than an introduction *sensu stricto*. Not included on the list are at

174least 33 other fishes that have dispersed across watersheds as a consequence of dam
175and reservoir construction that flooded a previously impenetrable barrier (*cf.* Júlio
176Júnior *et al.*, 2009). In the case of the African catfish *Clarias gariepinus*, Vitule *et al.*
177(2006) suggested their recording in Southern Brazil was a consequence of escapees
178from lakes used for recreational angling, although they have also been used in pond
179aquaculture since at least 1995 (FAO, 2011b). Compared with the native fish diversity
180of the country, the extent of the fish allodiversity resulting from aquaculture and
181angling appears relatively low, comprising of only six families (Table 1).

182

183**Freshwater fish and alien fish aquaculture production and value in Brazil**

184The reported total freshwater fish aquaculture production has been generally increased
185year-on-year in Brazil since recorded data were first available in 1995, with
186substantial increases in the late 1990s and early 2000s (Fig. 1). In the last 2 years for
187which data were available (2007 and 2008), total production was 209,812 and 210,906
188tonnes respectively. The production increases are reflected in their economic value
189(Fig. 1); in 2007 and 2008, the total value of freshwater fish aquaculture in Brazil was
190\$US 397 million and \$US 399 million respectively.

191

192 Of the species being cultured, the FAO statistics provide breakdowns for some
193species but amalgamate values for others. Of the alien fish that feature, the annual
194contribution of *C. carpio* to total aquaculture production between 1995 and 2008 has
195ranged from 17 to 46 %, and in value from 10 to 30 % (Fig. 1). Their annual
196production and economic value decreased through the 2000s. Data on the production
197and value of tilapia species in aquaculture have been combined for all the species,
198making it impossible to differentiate between alien and indigenous species. According

199to Vitule *et al.* (2009), however, *O. niloticus* are the principal tilapia used in Brazilian
200aquaculture. Thus, assuming that the majority of cultured tilapias are *O. niloticus*,
201then their contribution to total annual production and value has generally increased
202year-on-year since 1995 (Fig. 1). Their annual contribution to total aquaculture
203production between 1995 and 2008 ranged from 22 to 45 %, and in value from 30 to
20451 %. Of other alien fishes that are mentioned specifically in the statistics, *O. mykiss*,
205*C. gariepinus* and *I. punctatus* never contributed more than 6 % by production and 7
206% by value to the annual statistics in the period and their combined contributions in
2072008 were only 2.0 and 2.4 % respectively. Asian cyprinid species such as bighead
208carp *Hypophthalmichthys molitrix* and grass carp *Ctenopharyngodon idella* are
209utilised within Brazilian aquaculture but their specific production values are not
210provided and so are assumed to be insignificant (FAO, 2011b).

211

212 In combination, between 1995 and 2008, alien fishes contributed between 62 and
21380 % of total annual aquaculture production, and between 61 and 76 % of total annual
214value. Thus, freshwater fish aquaculture production in Brazil is highly dependent
215upon alien fishes.

216

217Value of recreational freshwater angling and the associated use of alien fishes in 218Brazil

219The practice of introducing alien fishes to enhance angling in impoundments has been
220a widespread activity in some areas of Brazil, such as the Upper Paraná basin where it
221was widespread until at least 1990 (Agostinho *et al.*, 2007); species such as *C. kelberi*
222and *C. piquiti* were introduced for the creation of sport fishing opportunities as both
223were valued for their fighting qualities (Espínola *et al.*, 2010). Although there are no

224available data relating to the economic value of using alien fishes in Brazilian
225fisheries, Holley *et al.* (2008) discussed the value of the indigenous peacock bass
226fishery of the Rio Negro River in Amazonia. Up to 1800 anglers are attracted to the
227region annually who spend up to \$US 6 million (Holley *et al.*, 2008). A study in the
228Brazilian Pantanal provided estimates of the value of recreational fishing to the
229region; the total annual welfare measure due to recreational fishing was estimated as
230ranging from \$US 35 to 56 million (Shrestha *et al.*, 2002). Although there was no
231suggestion this fishery was exploiting alien fishes, it does demonstrate that
232recreational anglers in Brazil are prepared to expend significant outlays on angling
233(Shrestha *et al.*, 2002). Thus, a successful fishery based on alien fishes may have the
234potential to generate significant economic income.

235

236Ecological risk assessments of alien fishes in Brazil

237According to Gozlan (2008), alien fishes that represent good candidates for use in
238freshwater aquaculture are those that generate considerable economic values but have
239minimum ecological consequences should they escape into the environment. In the
240case of alien fishes in Brazilian freshwater aquaculture, only *C. carpio* and *O.*
241*niloticus* generate substantial production and economic values (Fig. 1) but their FISK
242scores suggest a high degree of invasiveness (Fig. 2a). Their values of likelihoods of
243ecological impacts (Gozlan, 2010) suggest that whilst empirical evidence is less
244severe for *C. carpio*, for *O. niloticus*, invasion subsequent negative ecological impacts
245remain likely (Fig. 2b). Of the other alien fishes used in Brazilian aquaculture, *O.*
246*mykiss*, *I. punctatus* and *C. gariiepinus* may represent better candidates due to their
247lower ecological risks (Fig. 2a,b) but all are currently of low economic value.

248

249 In order to identify whether any other alien freshwater fishes being used across
250 South American aquaculture may be appropriate to use in Brazil on the basis of their
251 economic value versus ecological risk, data were collated for the continent from the
252 FAO database (Fig. 2). According to FISK scores, there are no alien fishes that
253 represented a strong candidate for use; all high economic value species were
254 associated with high scores (i.e. a high potential for invasiveness) (Fig. 2a). By
255 contrast, rainbow trout were shown to have a relatively low likelihood of ecological
256 impact (Gozlan, 2010) and generated significant economic values (Fig. 2b). However,
257 given their current low use in Brazil and their requirements of relatively cool, well
258 oxygenated water for production (FAO, 2011c), then it is unlikely there is scope for
259 any increase in production in Brazil due to the prevailing warm water conditions.

260

261 The analysis of comparing economic values against potential invasiveness and
262 likelihood of ecological impacts could not be completed for the *Cichla* species and
263 *Micropterus salmoides* used in fishery enhancement (Table 1). A FISK score was,
264 however, generated for the *Cichla* genus and the score (28.8) suggested a relatively
265 high risk of invasiveness and impact. Their likelihood of causing ecological impact
266 was also high, with no studies reporting negligible impact (*cf.* case study 1).
267 Regarding *M. salmoides*, their mean FISK score of 15.5 in Copp *et al.* (2009) suggests
268 their ability to invade is relatively low compared with *Cichla* fishes. Notwithstanding,
269 empirical evidence on their likelihood of causing deleterious impacts on native fishes
270 would appear high due to the high number of studies reporting negative consequences
271 arising from their invasion (e.g., Cambray & Stuart, 1985; Werner *et al.*, 1983;
272 Werner & Hall, 1988). For example, Gratwicke & Marshall (2001) reported a 99 %

273 population decline in fishes of the Cyprinidae family in farm dams in Zimbabwe when
274 *M. salmoides* was present.

275

276 Thus, it is apparent that the alien freshwater fishes most widely used in aquaculture
277 (*C. carpio*, *O. niloticus*) and recreational angling (*Cichla* species) in Brazil were
278 assessed as presenting considerable ecological risks, with high potential for invasive
279 populations developing. For *Cichla* species and *O. niloticus*, published empirical
280 studies on their introductions are reviewed to reveal whether actual impacts on native
281 fishes have been reported. For information on the ecological impacts of alien *C.*
282 *carpio*, see recent reviews by Weber & Brown (2009) and Cucherousset & Olden
283 (2011).

284

285 **Case study 1: Peacock basses (*Cichla* genus)**

286 Indigenous to Amazonia, peacock basses have been introduced across the Neotropical
287 Region where they have been described as an underestimated threat to native fish
288 communities (Pelicice & Agostinho, 2009). They are now widespread in many
289 Brazilian reservoirs, such as those in the Upper Paraná River basin, following
290 introductions for fishery creation (Espínola *et al.*, 2010). Introductions have generally
291 been successful in producing sustainable stocks for angling exploitation (Agostinho *et*
292 *al.*, 2007) and their populations tend to integrate successfully into the fish assemblage
293 (Paiva *et al.*, 1994; Oliveira *et al.*, 2006; Pelicice & Agostinho, 2009). The
294 invasiveness of *C. kelberi* is demonstrated by their subsequent wide dispersal
295 throughout the Upper Paraná River basin following initial introductions (Kullander &
296 Ferreira, 2006). Espínola *et al.* (2010) suggested it was the largest, deepest, most
297 transparent and warmest reservoirs that were most likely to be colonised. Data on

298their invasive genetics, diet, parasitology, reproductive traits and growth rates are now
299available in a number of studies from across their extended range in Brazil (e.g.,
300Oliveira *et al.*, 2006; Resende *et al.*, 2008; Martins *et al.*, 2009; Vieira *et al.*, 2009;
301Gomiero *et al.*, 2010a,b; Yamada *et al.*, 2011).

302

303Peacock basses have been described as exceptionally voracious predators and it is this
304that has raised concern over their impacts on native fishes (Santos *et al.*, 1994). Their
305piscivory has been described as seriously threatening native fish diversity (Godinho *et*
306*al.*, 1994), acting as a major force of biotic homogenisation (Latini & Petrere, 2004).
307Moreover, there is compelling evidence from empirical studies to support this. For
308example, in a study conducted in the Roasna Reservoir in the Paraná River basin, *C.*
309*kelberi* was introduced in 2004 (Pelicice & Agostinho, 2009). Comparison of samples
310collected before (2003) and after introduction (2007) revealed that in their presence,
311native fish diversity was significantly impacted; mean fish density reduced by 95 %
312and species richness by 100 %. There was the complete disappearance of many
313smaller fish species. It was predicted that complete assemblage extinction could occur
314by 2010. Introductions of *C. monoculus* into lakes in the River Doce basin also
315resulted in reduced native fish richness and diversity of the community (Latini &
316Petrere, 2004). Other studies have shown deleterious impacts by *C. kelberi* on prey
317fishes were not prevented by vegetation, with aquatic plants providing minimal
318refugia for prey fishes (Kovalenko *et al.*, 2010a). The vulnerability of native prey
319fishes to predation is also not due to a lack of recognition of *C. kelberi* as a predator,
320for anti-predator responses to a native fish predator and *C. kelberi* were similar
321(Kovalenko *et al.*, 2010b).

322 In summary, introductions of peacock bass into freshwaters waters across Brazil
323specifically and the Neotropical Region generally tend to result in invasions, with
324invasive populations then having substantial deleterious impacts on native fishes,
325affecting diversity and leading to biotic homogenisation.

326

327Case study 2: *Oreochromis niloticus*

328*Oreochromis niloticus* is commonly used in aquaculture across the world as the
329species represents a robust culture fish through their physiological tolerances and
330ability to withstand variable water chemistry, in conjunction with their production and
331husbandry being facilitated by their traits of multiple spawning, parental care, and
332high dietary plasticity (Attayde *et al.*, 2007; Freire & Prodocimo, 2007). Vitule *et al.*
333(2009) describe *O. niloticus* as the principal tilapia used in Brazilian aquaculture.
334Introductions into the environment are usually associated with releases from
335aquaculture facilities (Peterson *et al.*, 2005; Strecker, 2006; McCrary *et al.*, 2007) and
336following an introduction, should ambient conditions be suitable, then invasive
337populations can form rapidly (Weyl, 2006). Zambrano *et al.* (2006) suggested that
338suitable habitats for *O. niloticus* invasions were present across much of South
339America and particularly in Brazil.

340

341 In freshwater habitats where *O. niloticus* develops invasive populations, native
342species have been reported to be directly and adversely impacted through decreased
343abundance and even extinctions (Weyl, 2006). These result from overlaps in habitat
344and resource utilisation, and competition for spawning sites (Peterson *et al.*, 2004,
3452005; Canonico *et al.*, 2005; Vitule *et al.*, 2009). Indirect impacts on native fishes
346include modifications to habitat and water quality, including turbidity (Figueredo &

347Giani, 2005). Hybridisation with native *Oreochromis* species is also possible
348(Moralee *et al.*, 2000). Vitule *et al.* (2009) discussed the deleterious effect of *O.*
349*niloticus* on the native fish fauna of Brazil and described their adverse effects on the
350community structure of native fishes, their impacts in reducing the abundance of
351planktonic microcrustaceans and the lowering of water transparency through
352increased abundance of microalgae (Gurgel & Fernando, 1994; Attayde *et al.*, 2007;
353Okun *et al.*, 2008). The associated release of high quantities of nutrients into the
354environment from *O. niloticus* cage aquaculture systems has also been shown to result
355in local elevations in the availability of zooplankton which has been demonstrated to
356invoke changes in the feeding and diet of native zooplanktivorous fish (Strictar-
357Pereira *et al.*, 2010).

358

359 In summary, whilst the influence of *O. niloticus* on native fish diversity may not
360appear as profound as *Cichla* species, empirical evidence does also suggest substantial
361impacts on native fish abundance and richness, and indirect impacts on fish
362communities that include habitat modification and shifts in water quality.

363

364DISCUSSION

365

366Freshwater fish diversity in many river basins in Brazil has been adversely impacted
367by reservoir construction; substantial changes to habitats, blockage of migratory
368routes, disruptions to ecosystem functioning and major changes in native fish
369assemblages have been recorded and this is irrespective of alien fish introductions
370(Agostinho *et al.*, 2005, 2007, 2008; Júlio Júnior *et al.*, 2009; Gubiani *et al.*, 2010).
371However, following reservoir construction, high-value aquaculture and angling

372opportunities have been created in the new lentic habitats in which alien fishes feature
373prominently (Pelicice & Agostinho, 2009; Vitule *et al.*, 2009). Their use has thus
374provided substantial societal and economic benefits; for example, the total value of
375alien fishes in Brazilian freshwater aquaculture was in excess of \$US 250 million in
3762008 (FAO, 2011b). This benefit may come, however, with an ecological and
377conservation cost to native fishes. This is because the alien fishes most used in these
378activities were assessed as having traits that provide them with a high degree of
379invasiveness. Empirical studies validated this. Peacock basses have been described as
380a voracious predator acting as a major force of biotic homogenisation (*cf.* Case study
3811). The effect of Nile tilapia on native fishes was direct through competition and
382indirect through changes in water quality and habitat (*cf.* Case study 2). There are also
383numerous empirical studies and reviews available on negative impacts of invasive *C.*
384*carpio* populations (e.g., Weber & Brown, 2009; Cucherousset & Olden, 2011).

385

386 Thus, there is an apparent disparity between the societal and economic benefits
387accrued from using alien freshwater fishes in Brazilian aquaculture and angling, and
388the negative impacts they may incur on native fish diversity. As a consequence,
389identification of alternative management options to using these fishes is necessary that
390seek to reduce risk and impacts on native fishes but maintain benefits (Gozlan, 2008,
3912010). Regarding aquaculture, then consideration of using an increased proportion of
392indigenous fishes is already being considered through development of ‘technological
393packages’ for native fishes with economical potential, such as tambaqui (*Colosomas*
394species) and surubin (*Pseudoplatistoma* species) (FAO, 2011a). This is allied to Ross
395*et al.* (2008) who argued that if aquaculture is to continue to expand in developing
396markets whilst minimising negative impacts from using alien fishes, then it must

397investigate increasing the cultivation of indigenous species. The situation of using
398peacock basses in sport angling is more complex in that introductions have already
399resulted in invasions and so non-indigenous populations are already widespread.
400Preventing further introductions is, thus, unlikely to minimise impacts on native fish
401diversity in river basins where they are already present and dispersing. The spatial
402scale at which the fishes are now present may make any form of population control
403virtually impossible, given that the ability to manage invasive fish in the environment
404is inversely related to the spatial extent of their dispersal (Britton *et al.*, 2011b). To
405control and/ or eliminate peacock basses from invaded reservoirs may also require
406extensive ‘scorched-earth’ techniques such as piscicide application and this would be
407likely to incur substantial and unacceptable impacts on non-target native fishes
408(Britton *et al.*, 2011b).

409

410 As it is apparent that managing the alien fishes already present in Brazilian
411freshwaters would be technically demanding, highly expensive and questionable as to
412its ability to succeed (Britton *et al.*, 2011b), then their current status may have to be
413accepted. Management approaches could then focus on the implementation of risk
414assessment methodologies that aim to prevent introductions of new alien fishes and
415movements of non-indigenous fishes to areas outside of their existing ranges. Risk
416assessment procedures remain completely absent in decision making processes on
417alien fishes in Brazil (Agostinho *et al.*, 2005, 2007, 2008). Given the development of
418risk assessment methodologies in the last decade elsewhere in the world (e.g., Copp *et al.*,
4192005, 2009; Britton *et al.*, 2010, 2011a) then this can be remedied forthwith.
420Accordingly, to protect the native fish diversity of Brazilian freshwaters from the
421adverse effects of alien fish introductions, we recommend:

- 422 • Implementation of full risk assessment and risk management procedures when
423 proposed introductions of non-indigenous fishes are made for both aquaculture
424 and angling. Only species assessed as minimal risk to native fishes and their
425 diversity should be considered for introduction to open systems;
- 426 • No new introductions of piscivorous non-indigenous fishes into Brazilian
427 freshwaters for the creation and/ or enhancement of angling;
- 428 • Increased technological development of aquaculture systems for indigenous
429 fishes, with support provided to aquaculturists to switch production from alien
430 to indigenous fishes; and
- 431 • Formulation of consultative bodies comprising of all relevant stakeholders to
432 assist implementation of the above recommendations, with the provision of
433 information on the societal and economic benefits of using alien fishes in
434 aquaculture and angling versus their ecological risk.

435

436 In summary, Brazil is a country with substantial diversity in native fish
437 assemblages. This review has demonstrated that alien fishes do provide high societal
438 and economic benefits in their utilisation in aquaculture and recreational angling, but
439 this is currently to the detriment of native fish diversity and ecological integrity. Thus,
440 if the native fish diversity across Brazil is to be maintained, then policies require
441 implementing on alien fishes that seek to prevent their further biological
442 homogenisation. Given the lack of risk based tools for alien fishes across the
443 Neotropical Region more generally, then these policies will have application and
444 relevance at this wider spatial scale.

445

446

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448

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645 *niloticus*) in American freshwater systems. *Canadian Journal of Fisheries and*
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647

648 Table 1. Alien fishes recorded in Brazilian freshwaters, their introduction pathway
 649 and whether they were introduced from outside of Brazil or were translocated from
 650 within Brazil. Note the list does not include the movement of fish between watersheds
 651 occurring as a consequence of dam construction (*cf.* Júlio Júnior *et al.*, 2009).

Species	Family	Introduction Pathway	Introduction or translocation
<i>Micropterus salmoides</i>	Centrarchidae	Angling	I
<i>Cichla spp.</i> (including <i>C. piquiti</i> and <i>C. kiberi</i>)	Cichlidae	Angling	T
<i>Plagioscion squamosissimus</i>	Cichlidae	Angling	T
<i>Geophagus proximus</i>	Cichlidae	Aquaculture	T
<i>Ctenopharyngodon idella</i>	Cyprinidae	Aquaculture	I
<i>Cyprinus carpio</i>	Cyprinidae	Aquaculture	I
<i>Hypophthalmichthys molitrix</i>	Cyprinidae	Aquaculture	I
<i>Ictalurus punctatus</i>	Ictaluridae	Aquaculture	I
<i>Oncorhynchus mykiss</i>	Salmonidae	Aquaculture	I
<i>Oreochromis mossambicus</i>	Cichlidae	Aquaculture	I
<i>Oreochromis niloticus</i>	Cichlidae	Aquaculture	I
<i>Oreochromis urolepis hornorum</i>	Cichlidae	Aquaculture	I
<i>Tilapia rendalli</i>	Cichlidae	Aquaculture	I
<i>Clarias gariepinus</i>	Clariidae	Aquaculture/ angling	I
<i>Oreochromis aureus</i>	Cichlidae	Aquaculture	I
<i>Carassius auratus</i>	Cyprinidae	Other	I
<i>Xiphororus maculatus</i>	Poeciliidae	Other	I
<i>Lepomis cyanellus</i>	Centrarchidae	Other	I
<i>Apistogramma erythrura</i>	Cichlidae	Other	I
<i>Omobranchus punctatus</i>	Blenniidae	Other	I
<i>Betta splendens</i>	Osphronemidae	Other	I

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654 **Figure captions**

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656 Figure 1. (a) Total annual freshwater fish aquaculture production (○) and the annual
657 production of *Cyprinus carpio* (■) and tilapia species (□) in Brazil between 1995 and
658 2008. (b) Total annual freshwater fish aquaculture value (○) and the annual value of
659 *Cyprinus carpio* (■) and tilapia species (□) in Brazil between 1995 and 2008.

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661 Figure 2. (a) Economic value of freshwater fish aquaculture per alien species in South
662 America in 2008 versus the FISK scores of the species (Copp *et al.*, 2009; Birton *et*
663 *al.*, 2010). (b) Economic value of freshwater fish aquaculture per alien species in
664 South America in 2008 versus the likelihood of their ecological impact (Gozlan,
665 2010). Key: Cc *Cyprinus carpio*; Cg *Clarias gariepinus*; On m *Oncorhynchus*
666 *mykiss*; Ip *Ictalurus punctatus*; CI *Ctenopharyngodon idella*; Om *Oreochromis*
667 *mossambicus*; On *Oreochromis niloticus*; Ar *Acipenser ruthenus*; Ab *Acipenser*
668 *baeri*; St *Salmo trutta*; and Ss *Salmo salar*.

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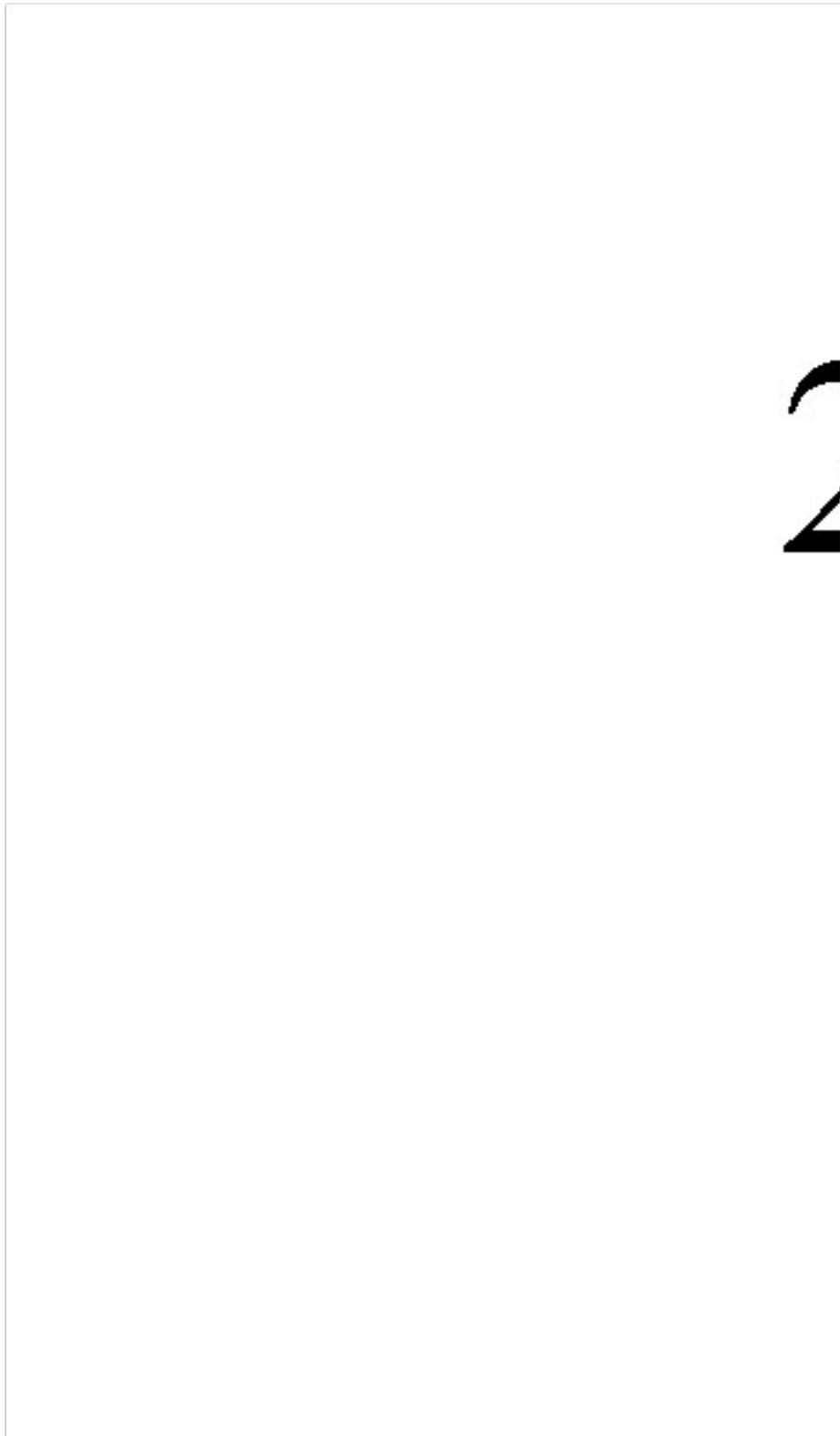
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693Figure 1.



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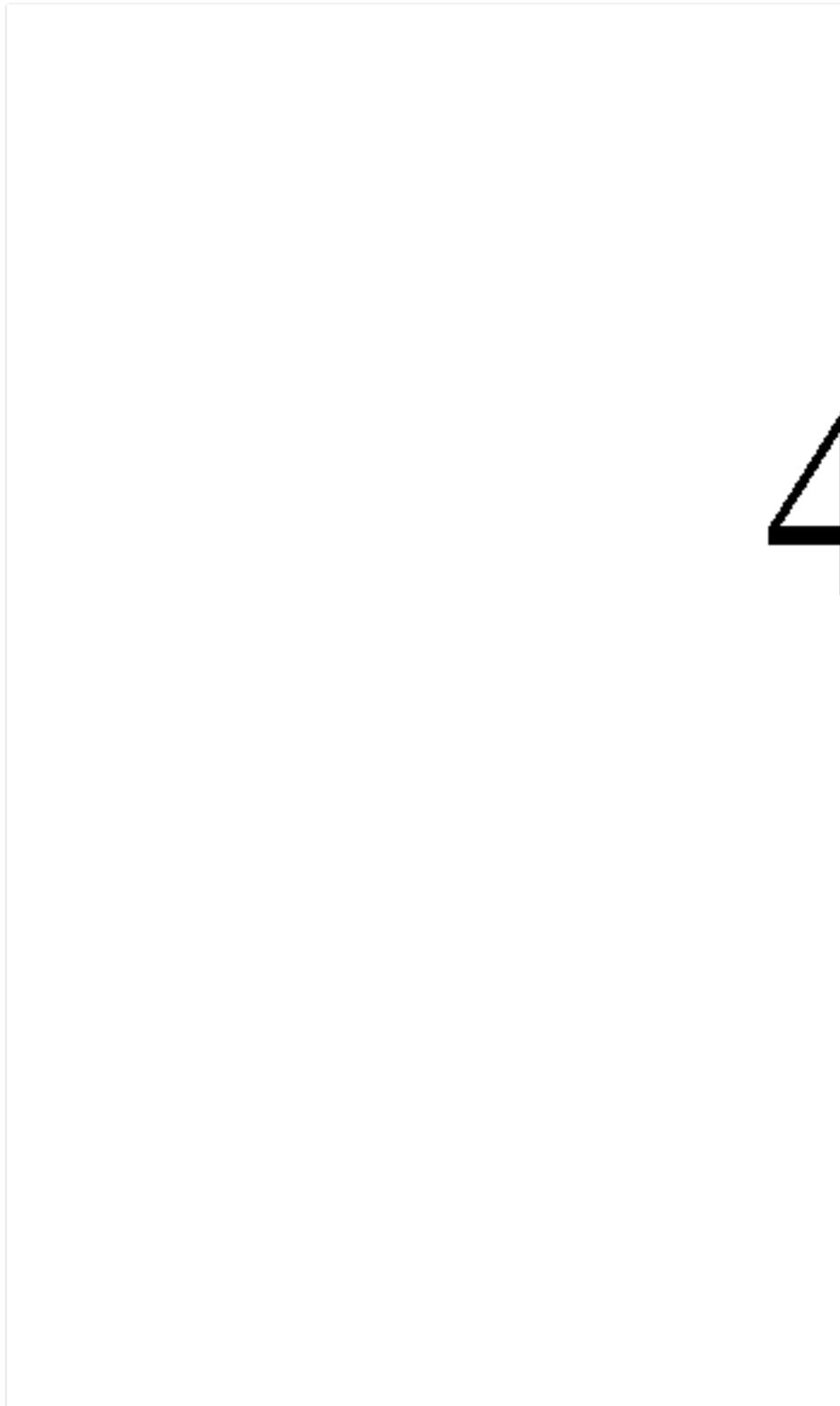
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718Figure 2.



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